



## Tansley review

# Competition between roots and microorganisms for nitrogen: mechanisms and ecological relevance

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### Summary

Demand of all living organisms on the same nutrients forms the basis for interspecific competition between plants and microorganisms in soils. This competition is especially strong in the rhizosphere. To evaluate competitive and mutualistic interactions between plants and microorganisms and to analyse ecological consequences of these interactions, we analysed 424 data pairs from 41 <sup>15</sup>N-labelling studies that investigated <sup>15</sup>N redistribution between roots and microorganisms. Calculated Michaelis–Menten kinetics based on  $K_m$  (Michaelis constant) and  $V_{max}$  (maximum uptake capacity) values from 77 studies on the uptake of nitrate, ammonia, and amino acids by roots and microorganisms clearly showed that, shortly after nitrogen (N) mobilization from soil organic matter and litter, microorganisms take up most N. Lower  $K_m$  values of microorganisms suggest that they are especially efficient at low N concentrations, but can also acquire more N at higher N concentrations ( $V_{max}$ ) compared with roots. Because of the unidirectional flow of nutrients from soil to roots, plants are the winners for N acquisition in the long run. Therefore, despite strong competition between roots and microorganisms for N, a temporal niche differentiation reflecting their generation times leads to mutualistic relationships in the rhizosphere. This temporal niche differentiation is highly relevant ecologically because it: protects ecosystems from N losses by leaching during periods of slow or no root uptake; continuously provides roots with available N according to plant demand; and contributes to the evolutionary development of mutualistic interactions between roots and microorganisms.

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**Key words:** carbon (C) and nitrogen (N) turnover, competition for nitrogen (N) and phosphorus (P), mutualism, niche differentiation, nutrient acquisition, plant–microbe interactions, priming effect, rhizosphere ecology.

## I. Introduction

All living organisms require nearly the same nutrients for their maintenance, growth, and reproduction. This forms the basis for interspecific competition between plants and microorganisms in nutrient-limited soils. This competition for available nutrients is particularly strong in the rhizosphere because of at least three factors.

First, plants and microorganisms in all soils (except soil types such as Phaeozems, Chernozems, and some Fluvisols) are limited by at least some nutrients, mainly nitrogen (N) and phosphorus (P) and certain other macro- and microelements. For soils in humid climates, this limitation is mainly ascribed to annual nutrient leaching (Lehmann & Schroth, 2003). Soils of semiarid and arid regions are not subjected to intensive chemical weathering, eliminating or reducing nutrient release from primary minerals. By contrast, in both tropical and wet subtropical soils that have very high weathering rates, only a few nutrients are present in the deeply weathered parent materials. Additionally, the remaining iron (Fe) and aluminium (Al) oxides bind nutrients occurring in anionic forms (e.g. P and molybdenum (Mo)) and render them unavailable for plants and microorganisms (Barber, 1996).

Secondly, continuous nutrient uptake by plants leads to the development of strong depletion zones around the roots (Fusseder & Kraus, 1986; cited after Jungk, 2001). Within 1–3 mm from the root surface, the concentration of nutrients that diffuse slowly, such as P, might decrease by *c.* 5–10 times compared with that in the nonrhizosphere soil (Tinker & Nye, 2000; Jones *et al.*, 2004). On the other hand, mobile nutrients in soil solution, such as NO<sub>3</sub><sup>-</sup>, do not develop strong depletion zones around roots; however, their concentration decreases extremely rapidly as a result of root uptake (Tinker & Nye, 2000). By contrast, NH<sub>4</sub><sup>+</sup> develops strong depletion zones around roots because it is not very mobile (Orcutt & Nilsen, 2000).

A third factor contributing to an increased competition is the release of large amounts of easily available carbon (C) into the rhizosphere (Kuzakov & Domanski, 2000; Nguyen, 2003). This available C leads to an increase in the abundance (Newman & Watson, 1977; Lynch, 1990; Kapoor & Mukerji, 2006; Saharan & Nehra, 2011), activity, and growth of microorganisms in the rhizosphere (Oger *et al.*, 2004; Blagodatskaya *et al.*, 2009, 2010), and consequently depletes the remaining available nutrients by microbial uptake and immobilization (Zak *et al.*, 2000).

Because of these three factors, particularly the last two, nutrient limitation for roots and microorganisms in the rhizosphere is far greater than that in the nonrhizosphere soil; this leads to strong competition between roots and microorganisms for nutrients.

## II. General solution: trade of C for nutrients

Despite this competition, microorganisms and roots depend on each other and have developed various mechanisms for symbiotic coexistence. One of the most well known, and probably the best investigated, mechanisms is the fixation of atmospheric N<sub>2</sub> by rhizobia and *Frankia*, which occurs on the basis of the C and energy provided by the roots of legumes and nonlegumes (Newton *et al.*, 2008; Franche *et al.*, 2009). The other is the development of mycorrhizal fungi on the root surface (ectomycorrhizas) or within the roots (e.g. ericoid and arbuscular mycorrhizas) with highly proliferating hyphae that penetrate into small pores and develop an all-embracing network (Read & Perez-Moreno, 2003; Southworth *et al.*, 2005; Simard *et al.*, 2012). This allows intensive acquisition of nutrients from a considerably larger soil volume. Additionally, mycorrhizal fungi substantially enhance N acquisition by plants from organic sources (Hodge & Fitter, 2010; Talbot & Treseder, 2010; Whiteside *et al.*, 2012). The two above-mentioned examples (rhizobacteria and mycorrhizas), however, are based not on the competition between roots and microorganisms, but on the mutualism evolved over millions of years of co-evolution (Lambers *et al.*, 2009; Kiers *et al.*, 2010). Roots provide microorganisms with C, and in turn obtain nutrients, because these microorganisms efficiently acquire nutrients from sources that are chemically or spatially unavailable to plants.

Numerous microorganisms in the soil are growing because of the absence of easily available substrates. The easily available C released by roots stimulates microbial growth in the rhizosphere, leading to the mining of additional N from soil organic matter (SOM; Kuzyakov, 2002; Luo *et al.*, 2006), that is, the production of some extracellular enzymes and thus enhancement of subsequent SOM decomposition (German *et al.*, 2011). This increases the N content in microbial biomass, and often also in plants (Hu *et al.*, 2006). A similar situation has been observed in many studies under field (Hamilton & Frank, 2001; Finzi *et al.*, 2007) and controlled conditions (De Graaff *et al.*, 2010; Medina-Roldán & Bardgett, 2011). Therefore, three effects (i.e. mutualism, neutralism, and competition; Table 1) can be observed in the interactions between plants and microorganisms. However, thus far, it is unclear how roots and microorganisms coexist and benefit from each other, although they depend on the same nutrients and strongly compete for them, particularly in the rhizosphere.

We describe these interactions between plants and microorganisms (Table 1) and focus mainly on their mutualism and competition. As N is the main limiting element in many terrestrial ecosystems (LeBauer & Treseder, 2008) and most published studies deal with N, this review mainly describes the processes

**Table 1** Interplay of positive (+), neutral (0) and negative (–) effects between plants and microorganisms (MO)

		Effects on plants		
		+	0	–
Effects on microorganisms	+	Mutualism	Commensalism	Parasitism on plant
	0	Commensalisms	Neutralism	Antagonism of plant
	–	Parasitism on MO	Antagonism of MO	Competition

related to N acquisition. Importantly, the mechanisms described for N might be similar for P and other nutrients. We hypothesize that N-cycling by microbes from SOM ultimately benefits plants, and that competitive and mutualistic interactions develop mainly on the basis of temporal and spatial niche differentiation for N acquisition when roots occupy new soil volume. To test this hypothesis, we conducted a survey of studies published since 1970, and here we suggested some of principles underlying the coexistence of roots and microorganisms in the rhizosphere.

### III. Methods

Evaluating mutualisms between roots and microorganisms and their competition for nutrients requires consideration of the following three important aspects: the spatio-temporal dynamics of available nutrients in the rhizosphere; the temporal dynamics of nutrient partitioning between microorganisms and plants; and the dynamics of the fitness and competitive abilities of microorganisms and plants. These dynamics can vary markedly depending on the prevalence of competition or mutualisms (Schimel & Bennett, 2004), and depend on the time that has elapsed since the start of N uptake. Accordingly, the most valuable studies include repeated measures of N in roots and microorganisms as well as changes in N availability in soil.

#### 1. Literature survey and data selection criteria

According to the three dynamics mentioned in the above paragraph, we selected the following parameters: Michaelis–Menten kinetics, that is, the maximal uptake rate ( $V_{\max}$ ) and substrate affinity ( $K_m$ ) for N uptake by microorganisms and roots, as well as  $^{15}\text{N}$  redistribution between plants and microorganisms. We searched data published since 1970 for this review in the *Web of Science*, *Google Scholar*, *CAB Abstracts*, and book series. The parameters obtained from these surveys were screened for explicit evaluation of the ecological significance of plant–microbe competition for N.

In all, we collected 118 articles for the meta-analysis reporting N uptake by both plant roots and microorganisms. Forty-two studies were screened for the  $V_{\max}$  and  $K_m$  parameters of N uptake by microorganisms (Supporting Information Table S1). These parameters were differentiated on the basis of the uptake kinetics of  $\text{NH}_4^+$ ,  $\text{NO}_3^-$ , and amino acids for microorganisms grown in culture or in soil. Another 35 studies were screened for the  $V_{\max}$  and  $K_m$  parameters of N uptake by roots (Table S2). Furthermore, 41  $^{15}\text{N}$ -labelling studies were collected to analyse  $^{15}\text{N}$  redistribution between plants and microorganisms (Table S3). Only those  $^{15}\text{N}$  studies that simultaneously investigated  $^{15}\text{N}$  uptake and partitioning between plants and microorganisms were included in this database.

#### 2. Data assembly

The first database focused on microorganisms and contained 249 pairs of  $V_{\max}$  and  $K_m$  values for N uptake from the 42 studies. The second database contained 436 pairs of  $V_{\max}$  and  $K_m$  values for N uptake by roots from 35 studies. All plant species were classified into three functional groups: grasses, forbs, and trees (grasses are

herbs, and forbs are herbs that are not grasses; Table S2). All  $V_{\max}$  values from original studies were recalculated and expressed as  $\mu\text{mol g}^{-1} \text{DW h}^{-1}$  (dry root weight used for plants and dry cell weight for microorganisms). When the values were presented for fresh weight, they were converted to dry weight using a coefficient of 0.2. All  $K_m$  values were presented in  $\mu\text{M}$ .

We used  $V_{\max}$  and  $K_m$  parameters from the first two databases to predict the results of the competition between plants and microorganisms for N uptake. We assumed that the Michaelis–Menten functions evaluate competition outcome in a broad range of  $\text{NO}_3^-$ ,  $\text{NH}_4^+$ , and amino acid concentrations, and possible chemical niche differentiation on the basis of microbial uptake of all three N forms in soil: reduced inorganic (ammonia), oxidized inorganic (nitrate), and reduced organic (amino acids). We evaluated  $V_{\max}$  and  $K_m$  for microorganisms and plants and summarized them separately for the three N forms: ammonia ( $\text{NH}_4^+$ , including studies on urea); nitrate ( $\text{NO}_3^-$ ); and amino-N, including mainly amino acids and also some amino sugar studies.

We assumed that these calculations overestimate the rate of plant N uptake under soil conditions because of the spatial distribution of roots in soil, as well as the considerably lower diffusion of nutrients in soil vs nutrient solution. However, to our knowledge, thus far, no approach has evaluated the extent of this overestimation.

#### 3. $^{15}\text{N}$ -labelling approach

The third database was developed to evaluate  $^{15}\text{N}$  partitioning between microorganisms and plants in order to represent their direct and ecologically relevant competition for N. This database contains 424 data pairs from 41  $^{15}\text{N}$ -labelling studies (Table S3) that investigated  $^{15}\text{N}$  redistribution between plants and microorganisms across different ecosystem types (Table 2).

Adding very small amounts of  $^{15}\text{N}$  label allows short-term evaluation of competition. Such short-term experiments reflect dynamic N uptake by roots and microorganisms, and therefore the competition in the rhizosphere over a specific growth period. Commonly,  $^{15}\text{N}$  is added in mineral ( $\text{NH}_4^+$  or  $\text{NO}_3^-$ ) or organic (amino acids) forms. The amounts of added  $^{15}\text{N}$  are usually very small (Biernath *et al.*, 2008; Xu *et al.*, 2011a) and, therefore, the equilibrium between the N forms remains unaffected and  $^{15}\text{N}$  functions as a label rather than an N fertilizer. The  $^{15}\text{N}$  label is added on the soil surface (Xu *et al.*, 2003; Littschwager *et al.*, 2010) or injected directly at a specific soil depth (Jackson *et al.*, 1989; Hart *et al.*, 1993; McKane *et al.*, 2002). This approach ensures that plants and microorganisms have equal access to the added  $^{15}\text{N}$ .

We used the third database to evaluate the plant–microbial competition for N. This evaluation was made on the basis of the ratio of  $^{15}\text{N}$  in microbial biomass (analysed by the extraction–fumigation approach) and in plants (including roots). Ratios of  $> 1$  show that microorganisms outcompete roots. Additionally, the dynamics of the  $^{15}\text{N}$  ratio in plants and microorganisms is important, because it shows the temporal changes in the acquisition of the limiting resource (Magyar *et al.*, 2007).

Most studies on root–microorganism competition have been conducted using this  $^{15}\text{N}$ -labelling approach. We therefore

**Table 2** Ecosystem types for the 41 <sup>15</sup>N-labelling studies used in this study

No.	Subtype	Vegetation	References
1	Alpine/subarctic	Shrub	Andresen <i>et al.</i> (2008)
2	<i>Calluna</i> /grasses/mosses	Heath	Andresen <i>et al.</i> (2009)
3	<i>Calluna</i> /grasses/mosses	Heath	Andresen <i>et al.</i> (2011)
4	<i>Acomastylis rossii</i>	Tundra	Ashton <i>et al.</i> (2008)
5	Temperate grassland	Grassland	Bardgett <i>et al.</i> (2003)
6	<i>Holcus lanatus</i>	Mesocomis	Barnard <i>et al.</i> (2006)
7	<i>Fraxinus excelsior</i> + <i>Dactylis glomerata</i>	Forest	Bloor <i>et al.</i> (2009)
8	<i>Quercus douglasii</i>	Forest	Cheng & Bledsoe (2004)
9	Blue oak ( <i>Quercus douglasii</i> )	Forest	Cheng & Bledsoe (2005)
10	Shortgrass prairie	Grassland	Clark (1977)
11	Heath	Tundra	Clemmensen <i>et al.</i> (2008)
12	<i>Festuca rubra</i>	Grassland	Dunn <i>et al.</i> (2006)
13	Birch	Forest	Grogan & Jonasson (2003)
14	<i>Festuca</i> / <i>Agrostis</i> / <i>Gallium</i>	Grassland	Harrison <i>et al.</i> (2007)
15	Temperate grassland	Grassland	Harrison <i>et al.</i> (2008)
16	Annual grassland	Grassland	Hart <i>et al.</i> (1993)
17	<i>Puccinellia phryganodes</i>	Mash	Henry & Jefferies (2003a)
18	<i>Puccinellia phryganodes</i>	Mash	Henry & Jefferies (2003b)
19	<i>Lolium perenne</i>	Crop	Hodge <i>et al.</i> (2000b)
20	Barley	Crop	Inselsbacher <i>et al.</i> (2010)
21	Grassland	Grassland	Jackson <i>et al.</i> (1989)
22	Alpine moist meadow	Grassland	Jaeger <i>et al.</i> (1999)
23	<i>Atriplex parryi</i>	Shrub	James & Richards (2006)
24	<i>Carex</i>	Grassland	Kastovska & Santruckova (2011)
25	<i>Kobresia myosuroides</i>	Tundra	Lipson & Monson (1998)
26	Dry alpine meadow	Grassland	Lipson <i>et al.</i> (1999)
27	White spruce	Forest	McFarland <i>et al.</i> (2010)
28	<i>Calluna</i> /grasses/mosses	Heath	Miller <i>et al.</i> (2009)
29	Grass-legume	Crop	Nannipieri <i>et al.</i> (1985)
30	Acidic site	Tundra	Nordin <i>et al.</i> (2004)
31	Tussock	Tundra	Schimel & Chapin (1996)
32	Subalpine	Heath	Sorensen <i>et al.</i> (2008b)
33	Alpine/subarctic	Tundra	Sorensen <i>et al.</i> (2008a)
34	Tropical montane	Forest	Templer <i>et al.</i> (2008)
35	Pine clearcut	Shrub	Vitousek & Matson (1984)
36	Steppe	Grassland	Wu <i>et al.</i> (2011)
37	<i>Kobresia pygmaea</i>	Grassland	Xu <i>et al.</i> (2003)
38	<i>Kobresia pygmaea</i>	Grassland	Xu <i>et al.</i> (2004)
39	<i>Kobresia humilis</i>	Grassland	Xu <i>et al.</i> (2011b)
40	<i>Allium tricoccum</i>	Forest	Zak <i>et al.</i> (1990)
41	Northern hardwood forest	Forest	Zogg <i>et al.</i> (2000)

evaluated the following parameters of the competition: (1) <sup>15</sup>N allocation in plants vs <sup>15</sup>N allocation in microorganisms, and (2) the dynamics of <sup>15</sup>N distribution between plants and microorganisms over time.

#### IV. Interactions between roots and rhizosphere microorganisms

##### 1. Sequence of processes for nutrient acquisition by occupation of new soil volume

The N uptake by roots is not even throughout the rooted soil volume and along the roots. The main uptake occurs in the root hair

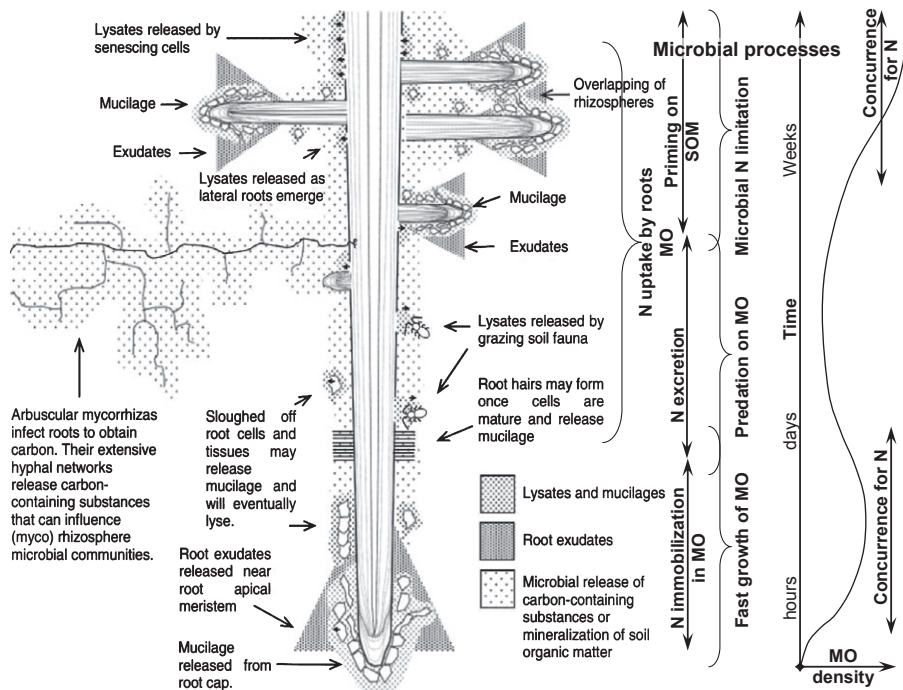
zone (Waisel *et al.*, 2002), *c.* 1–5 cm behind the root tip (Fig. 1). The zones behind the root hairs are responsible for only minimal nutrient uptake and are not considered here. Therefore, when a new soil volume is occupied, the duration of N uptake in the individual zones is limited by the passage in the root from the root tip to the root hair zone and just behind it (Clarkson, 1985; Ingestad & Ågren, 1988). The distance from the root tip to the end of the root hair zone is *c.* 1–5 cm, and roots grow from 0 to 2 cm d<sup>-1</sup>. This limits the period of direct N uptake from newly occupied soil to a few days (Thaler & Pagès, 1998).

Generally, root hairs have a life span of a few days, although they may persist for longer periods, particularly in grasses (Mengel *et al.*, 2001; Gregory, 2006). During root growth, different root zones release rhizodeposits of different compositions (Fig. 1). Although roots might take up some of the released organics (mainly low-molecular-weight organics; Näsholm *et al.*, 1998, 2009; Kuzyakov & Jones, 2006; Biernath *et al.*, 2008; Hill *et al.*, 2012), a substantially greater release than uptake results in a net release. When root hairs die and lyse, this C also contributes to rhizodeposition. Most of the C released as rhizodeposition is readily available, and microorganisms use this C for growth. Microbial utilization of rhizodeposits is very fast and occurs within a few hours for exudates (Jones & Kielland, 2002; Jones *et al.*, 2005; Kuzyakov & Jones, 2006; Kemmitt *et al.*, 2008; Fischer & Kuzyakov, 2010; Fischer *et al.*, 2010) and a few days for sloughed-off cell walls and root hairs (Dormaar, 1992).

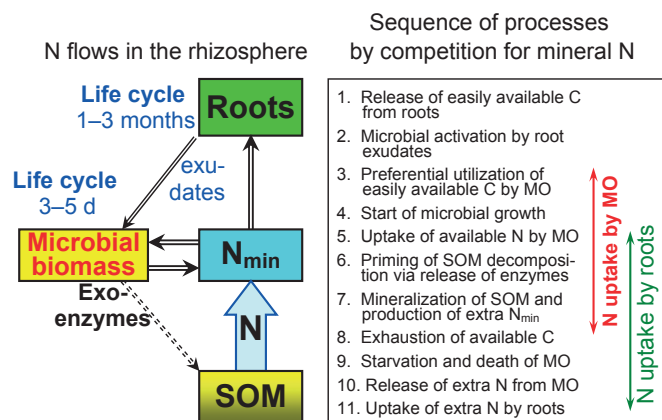
Considering these short utilization periods and the turnover time of rhizosphere microorganisms ranging from days to weeks (Staddon *et al.*, 2003; Schmidt *et al.*, 2007; Blagodatskaya *et al.*, 2011), we assume that the excess of easily available C is depleted within a few days via microbial uptake, utilization, and decomposition. This causes the microorganisms that were previously growing on the excess of substrate to starve. This absence of new C input and continuous consumption of C in the microorganisms leads to the release of N immobilized in microbial biomass into the soil. This process sequence results (Fig. 2) in the availability of N for plants.

As described in the sequence of processes involving the competition between plants and microorganisms for available N, the microorganisms stimulated by available C begin capturing N earlier than the roots (vertical arrow, right, in Fig. 2). This leads to a temporary decrease in the available N for the roots, and this N limitation in turn stimulates them to release additional C (Merckx *et al.*, 1987; Liljeroth *et al.*, 1990; Kuzyakov *et al.*, 2001). Although the mechanisms underlying the stimulation of C release by roots as a result of N limitation remain unclear, the increase in C is associated with the development of more abundant fine roots, and consequently higher rhizodeposition. Clarifying the C release by roots as a result of N limitation could significantly contribute to our understanding of mutualistic mechanisms, that is, whether they occur at the level of the root system or individual roots.

Following the uptake of root exudates, microorganisms use the C for growth and maintenance respiration. As N is limited even for microorganisms, they produce certain extracellular enzymes for the mineralization of poorly available C sources, such as SOM, to obtain N (Schimel & Weintraub, 2003; Manzoni *et al.*, 2012). These depolymerizing extracellular enzymes are glycosidases,



**Fig. 1** Microbial processes along the growing root and released rhizodeposits. The competition for nitrogen (N) between roots and microorganisms peaks at periods of strong microbial growth. Strong root–microorganism competition for N as well as stimulation of microbial activity by easily available rhizodeposits accelerates the decomposition of soil organic matter (SOM) for additional N mineralization (real priming effect). The time line on the right represents the period after the root occupies new soil volume. The microorganism (MO) density is presented as changes compared with root-free soil. Compilation from Kuzyakov (2002) and Dennis *et al.* (2010).



**Fig. 2** Sequence of processes when ingrowing roots occupy new soil volume and during interactions between roots and microorganisms for carbon (C) and nitrogen (N) uptake. Despite the initial uptake of N mineralized from soil organic matter (SOM) by microorganisms (MO), their much shorter life cycle compared with that of plant roots leads to the release of acquired N back into the soil; this mineral N is then available for root uptake.  $N_{min}$ , mineralized N. See text for further explanation.

phenoloxidas, peroxidases, glucosaminidas, and peptidas (Blagodatskaya & Kuzyakov, 2008), which break down polymers and generate dimers and monomers that are soluble and available for microbial uptake (Hill *et al.*, 2012). Because microbial turnover in the rhizosphere is very rapid (a few days), the C losses by respiration are very high. Consequently, these microorganisms become energy-starved and obtain their energy from amino acids. Microorganisms strip off the N from the amino acids, thereby making the C skeletons available for the tricarboxylic acid (TCA) cycle for energy and growth. Finally,  $NH_4^+$  is released as a metabolite in this process (ammonification). This form of N can be

absorbed by roots. Importantly, the net N release from starving or dying microorganisms is higher than the initially available N content because of microbial mining for extra N from SOM during root growth. Clarholm (1985) suggested another mechanism of N release from microorganisms that involves microbial grazing by soil animals, and which is particularly important in the rhizosphere (described in detail by Bonkowski *et al.*, 2009).

## 2. Effect of time on the nutrient flow direction

Two factors contribute to the redistribution of N after its initial acquisition by microorganisms and roots (Figs 1, 2): N flow direction, and life cycle duration of rhizosphere microorganisms and plants. The N flows driven by competition between roots and microorganisms are directed to the roots (Fig. 2). Decomposition of SOM by some exoenzymes releases minerals and amino acids and some amino sugars, which can be acquired by microorganisms and, after mineralization to  $NH_4^+$  or  $NO_3^-$ , by roots. Amino acids can also be used directly by roots, although compared with microorganisms, plant roots are often not very efficient competitors for this N source. As shown in Fig. 2, at the initial stage, microorganisms outcompete roots for inorganic N because of rapid growth rates and high surface-area-to-volume ratios compared with those of root hairs (Rosswall, 1982). Thus, even in soils with a high density of roots, for example in the upper 5 to 10 cm in grassland, most N will be allocated to microorganisms shortly after N addition or release from decomposing litter. This reduces N leaching losses as a result of the limited uptake capacity of roots.

At this stage, the second factor responsible for redistribution of N between microorganisms and roots arises: the duration of the life cycle of rhizosphere microorganisms and roots. The turnover time of the microorganisms is very short (a few days; Schmidt *et al.*, 2007) because of high C losses by respiration. In microorganisms,

the C released as CO<sub>2</sub> is lost, and their C:N ratio decreases. Despite this, the microbial C:N ratio is stable and ranges between 5 and 10 (average, *c.* 8; Cleveland & Liptzin, 2007). Mineral N is thus again available for microorganisms and roots. This cycle of the uptake and release of mineral N (and amino acid N) occurs within a few days according to the life cycle of the rhizosphere microorganisms.

Unlike the N flow in microorganisms, the net N flow to and in the roots is unidirectional. Despite some evidence for N rhizodeposition (Wichern *et al.*, 2008), the net N flow is directed to the roots, except in symbiotic diazotrophs. Although, in diazotrophy, the net flow of the sum of N species is directed to the roots, the minor efflux of the combined N obtained from the plant is offset by the influx of N generated during the fixation of N<sub>2</sub> by nodule bacteria. Therefore, roots acquire N in small portions, but more continuously and mainly unidirectionally. This leads to the continuous accumulation of N in roots and its depletion in the rhizosphere. Some of the N is involved in the synthesis of amino acids and proteins in roots, whereas another part is translocated to the shoots for further utilization. This indicates that, over long periods, plants acquire increasing amounts of N, which had already passed through a microbial cycle(s) and was initially stored in the SOM. At an ecosystem level, plants and microorganisms obtain the N as required (Kaye & Hart, 1997).

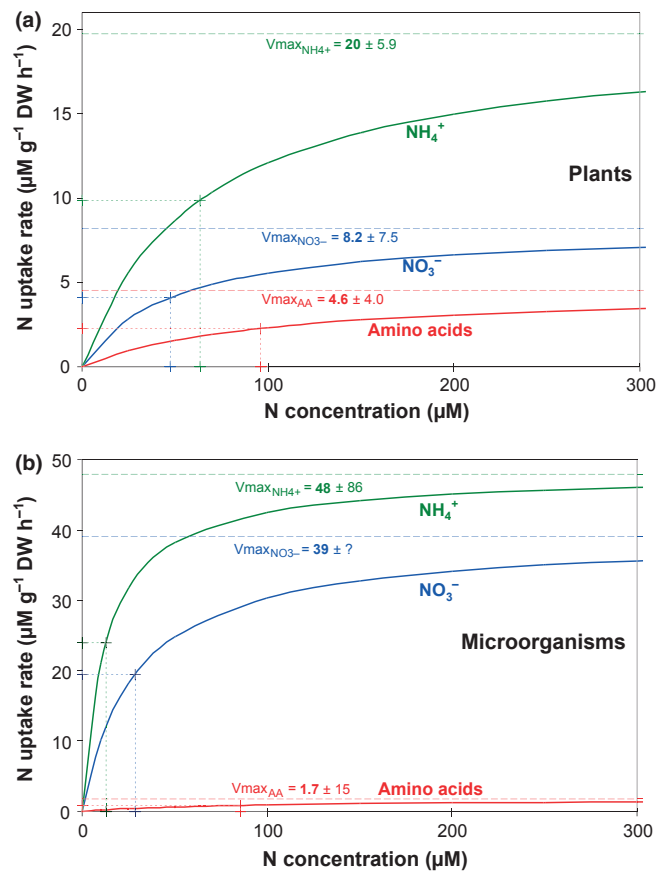
### V. Potential of plants and microorganisms for N uptake by their competition

Using data collected from the literature, we calculated medians of *K<sub>m</sub>* and *V<sub>max</sub>* values separately for three groups of N species (NO<sub>3</sub><sup>-</sup>, NH<sub>4</sub><sup>+</sup>, and amino acids) for roots and microorganisms (Fig. 3, Tables 3, S1, S2). Despite the high variation in *K<sub>m</sub>* and *V<sub>max</sub>* values, we used these values to calculate the uptake depending on N concentration. The Michaelis–Menten uptake kinetics clearly showed that microorganisms have a higher capacity (*V<sub>max</sub>*) for NO<sub>3</sub><sup>-</sup> and NH<sub>4</sub><sup>+</sup> uptake compared with roots, across the whole concentration range. This advantage is particularly pronounced at low concentrations of mineral N (< 50–100 μM), as a small *K<sub>m</sub>* indicates high affinity for NO<sub>3</sub><sup>-</sup> and NH<sub>4</sub><sup>+</sup> (Table 3).

**Table 3** Parameters of Michaelis–Menten kinetics: *K<sub>m</sub>* (Michaelis constant) and *V<sub>max</sub>* (maximum uptake capacity) for uptake of ammonia (NH<sub>4</sub><sup>+</sup>), nitrate (NO<sub>3</sub><sup>-</sup>) and amino acids by plants and microorganisms

N species	<i>K<sub>m</sub></i> (μM)				<i>V<sub>max</sub></i> (μM g <sup>-1</sup> h <sup>-1</sup> )			
	Mean	± SE	Median	No.	Mean	± SE	Median	No.
<b>Plants</b>								
NH <sub>4</sub> <sup>+</sup>	289	114	64	68	37	5.9	4.9	68
NO <sub>3</sub> <sup>-</sup>	79	11	48	60	37	7.5	8.2	64
Amino acids	265	105	96	82	15	4.0	4.6	82
<b>Microorganisms</b>								
NH <sub>4</sub> <sup>+</sup>	48	20	13	59	188	86	48	7
NO <sub>3</sub> <sup>-</sup>	377	227	29	20	39	–	39	1
Amino acids	530	111	86	86	52	15	1.7	75

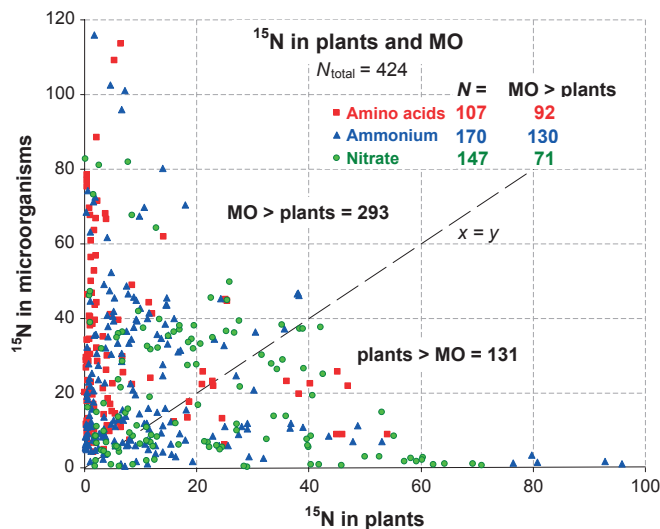
'No.' indicates the number of studies. All parameters from individual studies were recalculated for the same units.



**Fig. 3** Michaelis–Menten kinetics of nitrogen (N) uptake from nitrate (NO<sub>3</sub><sup>-</sup>), ammonia (NH<sub>3</sub>), and various amino acids (AA) by plants (a) and microorganisms (b). The lines were calculated on the basis of the medians of *K<sub>m</sub>* (Michaelis constant) and *V<sub>max</sub>* (maximum uptake capacity) values calculated from the database of studies with plants (35 studies; 436 *K<sub>m</sub>* and *V<sub>max</sub>* data points; see Supporting Information Table S2 for details) and microorganisms (42 studies; 249 *K<sub>m</sub>* and *V<sub>max</sub>* data points; see Table S1 for details). All *V<sub>max</sub>* values are expressed as μmol g<sup>-1</sup> DW h<sup>-1</sup> (dry root weight used for plants and dry cell weight for microorganisms). The mean ± 1 SD of *V<sub>max</sub>* values for NH<sub>4</sub><sup>+</sup>, NO<sub>3</sub><sup>-</sup>, and AA uptake by plants and microorganisms is also presented in the figure. The Michaelis–Menten kinetics clearly show that, within the N concentration range, but particularly at the low concentrations common in soils, NO<sub>3</sub><sup>-</sup> and NH<sub>4</sub><sup>+</sup> will be taken up by microorganisms considerably more rapidly than by plants.

The analysis was conducted assuming that nearly all root uptake studies were performed using nutrient solutions. If this was the case, the N diffusion rate would be considerably higher than that under normal soil conditions. Considering a more even distribution of microorganisms in soil vs roots, we expected that the advantages of microorganisms are even higher under real soil conditions compared with those afforded by the theoretical curves (Fig. 3) estimated on the basis of nutrient solution studies.

The hypothesis regarding uptake rate was partly confirmed while summarizing the <sup>15</sup>N studies (Fig. 4). We related <sup>15</sup>N in plants to that in microorganisms after a pulse addition of <sup>15</sup>N as NO<sub>3</sub><sup>-</sup>, NH<sub>4</sub><sup>+</sup>, or amino acids to soil. Confirming our theory, most of the <sup>15</sup>N was allocated to microorganisms (there are nearly twice the number of points above the *x* = *y* line in Fig. 4). Nonetheless, the red points in Fig. 4 clearly show a strong preference of



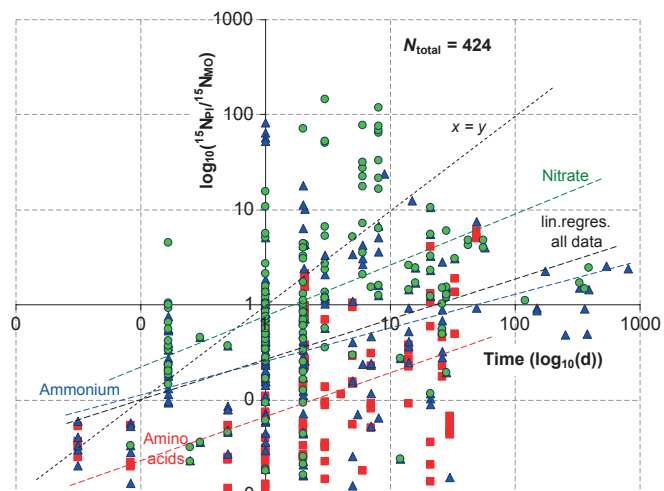
**Fig. 4** Review of  $^{15}\text{N}$  uptake by plants and microorganisms (MO) after  $^{15}\text{N}$  pulse addition to soil.  $^{15}\text{N}$  for plants and microorganisms is presented in the same mass units: in most studies, percentage of  $^{15}\text{N}$  input and, in some studies,  $\mu\text{g } ^{15}\text{N g}^{-1}$  soil. There are nearly twice as many values above the  $x = y$  line, suggesting that most of the added  $^{15}\text{N}$  is taken up by microorganisms. Note the clear differentiation between the three N forms: ammonia is mainly allocated to plants, and amino acids are taken up mainly by microorganisms. Nitrate is allocated preferentially to plants. The Kruskal–Wallis one-way analysis-of-variance-by-ranks test was used to examine the difference among ammonium, nitrate, and amino acids. A significant difference among these values was observed at  $P < 0.05$  (Dunn's method).

microorganisms for uptake of amino acids. This was not very obvious from Fig. 3, which shows  $K_m$  and  $V_{\max}$  values, but clearly confirms our expectation that, under soil conditions, the competition will be shifted towards microorganisms. Accordingly, most of the amino acids appearing in the soil will be recovered in microorganisms and not in plants. This result partly confirms our previous suggestion that amino acid uptake is of minor importance for the N nutrition of plants in some ecosystems (Biernath *et al.*, 2008; Xu *et al.*, 2008; Rasmussen *et al.*, 2010).

The Michaelis–Menten kinetics (Fig. 3) reflect the uptake of N, but do not consider subsequent redistribution of N between roots and microorganisms after the uptake. When the level of  $^{15}\text{N}$  in plants is compared graphically with that in microorganisms, the N dynamics within the system are still not clear (Fig. 4). We assumed that the life cycle of microorganisms and roots may have an effect on the later redistribution of N (see section VII). We therefore related the ratio between  $^{15}\text{N}$  in plants and that in microorganisms to the time after the addition of  $^{15}\text{N}$  to the soil (Fig. 5). This ratio clearly increased over time, indicating that, after initial preferential uptake by microorganisms, there is a redistribution leading to the relocation of N to plants.

## VI. Ecological relevance of competition between roots and microorganisms for N

Previous studies have not focused on *why* the competition between roots and microorganisms for N is important and *what* are the ecological consequences of this competition (Kaye &



**Fig. 5** Change in  $^{15}\text{N}$  distribution between plants and soil microorganisms over time after  $^{15}\text{N}$  input in soil.  $^{15}\text{N}$  distribution is presented as the ratio between  $^{15}\text{N}$  in plants and  $^{15}\text{N}$  in microorganisms ( $^{15}\text{N}_{\text{PL}}/^{15}\text{N}_{\text{MO}}$ ). Because of the enormous range of  $^{15}\text{N}_{\text{PL}}/^{15}\text{N}_{\text{MO}}$  and the duration of various experiments (from  $< 1$  h up to 3 yr), both variables are presented as decadal logarithms. The trend line (dashed lines) clearly shows that the ratio  $^{15}\text{N}_{\text{PL}}/^{15}\text{N}_{\text{MO}}$  increases over time and, consequently, more  $^{15}\text{N}$  will be relocated into the plants from microorganisms. For references see Table S3.

Hart, 1997; Hodge *et al.*, 2000a). In our view, this competition and the fast uptake of N by microorganisms have at least three important consequences for ecosystem development and species evolution.

First, most N cycling in soils is through organic forms (Wanek *et al.*, 2010; Inselsbacher & Näsholm, 2012) and most plants have the capacity to take up organic N in the form of free amino acids (Näsholm *et al.*, 2009). Therefore, mineralization of SOM provides available N, including mineral N (i.e.  $\text{NH}_4^+$  and  $\text{NO}_3^-$ ) and amino acid-N, for plants and microorganisms (Schimel & Bennett, 2004). Very rapid trapping of mineral N and amino acids by microorganisms (Jones & Kielland, 2002; Fischer *et al.*, 2010) is ascribed to their inherent uptake capacity (i.e. high surface-area-to-volume ratios and  $V_{\max}$ ). This is an important adaptation of ecosystems against possible N losses by leaching. Allocation of N within the microorganisms (temporal immobilization) protects it against leaching. Therefore, in those natural ecosystems in which microorganisms retain nearly all N, the leaching losses are minimal, even in cases where the plants are unable to take up all N after its release by litter decomposition. This is partly confirmed by the fact that, in strongly N-limited ecosystems such as alpine grasslands, microorganisms rapidly trap N (Song *et al.*, 2007).

In contrast to natural ecosystems, areas with intensive agriculture have an excess of mineral N added as fertilizers that cannot be trapped by microorganisms because of the absence of available C, leading to a considerably lower microbial biomass and activity. Normally, agricultural soils contain  $< 70$ – $50\%$  of SOM compared with virgin soils (Lal, 2003). This decrease mainly reflects reduced pools of easily available SOM (Six *et al.*, 2002; Chen *et al.*, 2009). The inability of microorganisms to trap and retain all fertilizer N leads to considerably higher N losses by leaching in agricultural soils (Havlin *et al.*, 1999; Kirchmann *et al.*, 2002) and certain other

disturbed ecosystems compared with natural ecosystems. Therefore, high N losses from agroecosystems are associated not only with abundant mineral fertilizers (the annual N turnover in many natural ecosystems is even higher than in croplands) but also with the particular inability of microorganisms to retain excess mineral N.

Secondly, this fast preliminary microbial trapping of mineral N and amino acids probably affects plant species competition. Plants having close mutualistic relationships with rhizosphere microorganisms can store the temporary excess N in the microorganisms. Therefore, the excess N will not be leached (see the second paragraph of this section) and lost to other species, but will be temporarily secured in microbial biomass close to the root surface. This offers such plant species advantages for N (and probably other nutrient) acquisition. In particular, enhanced rhizodeposition provides a strong selection pressure for those microorganisms that are retained close to the root surfaces and promotes evolutionary selection for mutualistic interactions.

The third important ecological consequence is the continuous provision of mineral N for roots at the time of maximal plant demand. In temperate ecosystems, the input of litter in autumn is temporarily decoupled from the plant demand in spring. This decoupling ranges in duration from a few months (e.g. for alpine grasslands, only the unfrozen soil period is considered) to approximately half a year (the winter period in humid temperate forests and grasslands). In many ecosystems (boreal and deciduous forests, and Mediterranean ecosystems), the period in which plant growth ceases is often characterized by high precipitation. If no N is retained by microorganisms, natural ecosystems would experience high leaching losses. The N released by the decomposition of plant residues is initially stored in microorganisms until the following spring, providing protection against leaching. The soil temperature increase in spring accelerates microbial turnover, and N stored in microorganisms over winter (or the cold period) is released in spring and becomes available for plant growth. At this time, the plants have high photosynthetic and growth rates and a maximal demand for N. A noteworthy fact is that not only photosynthesis and plant growth rate but also soil temperature depend on solar radiation. Therefore, the N release by microorganisms, driven by soil temperature, corresponds to maximal plant growth driven by photosynthesis. Both processes depend on solar radiation, but there is a lag in maximum soil temperature and maximum irradiance.

On the basis of the ecological significance of fast N uptake by microorganisms and its subsequent slow release when plants need it most, we suggest that the competition between plants and microorganisms for N (and probably some other nutrients) prevents leaching losses. This is actually a strategy to maintain ecosystem stability from an evolutionary perspective. The C demand by soil microorganisms is met by plant photosynthesis. To stabilize and even increase plant growth (and thus the amount of C captured by photosynthesis), microorganisms use this C to mobilize nutrients and, after a delay, provide a part of this for plants. Accordingly, the plant–microorganism competition for nutrients is facilitated at the rhizosphere at the ecosystem scale. This phenomenon effectively converts ‘competition’ to cooperation, particularly in N-limited ecosystems.

**Table 4** Effects of biotic and abiotic factors on competition between roots and microorganisms for nitrogen (N)

	N in plants	N in microorganisms	Competition	References
<b>Biotic factors</b>				
Root density ↑	↑↑↑	↓	↑↑	Cheng & Bledsoe (2004)
Mycorrhiza ↑	↑↑↑	↓	↑	Hodge <i>et al.</i> (2000a)
Root nodules ↑	↑	↑	↓	Kiers <i>et al.</i> (2003)
Photosynthesis ↑	↑	↓	↑	Xu <i>et al.</i> (2008)
Plant growth ↑	↑	↓	↑	
<b>Abiotic factors</b>				
Soil depth ↑	↓	↑	↓	Xu <i>et al.</i> (2011b)
Temperature ↑	↑	↓	↑	
N fertilization ↑	↑	↑	↓	Bloor <i>et al.</i> (2009)
pH of soil ↑	↓	↑	?	
<b>Ecosystem factors</b>				
NPP ↑	↑	↑	?	

The table can be read as follows. For ‘Root density’, for example: if root density increases (↑), then plant uptake of N increases (↑), N in soil decreases (↓), and the root–microorganism competition increases (↑). A question mark indicates uncertain roles where further investigation is required. NPP, net primary production.

## VII. Effect of biotic and abiotic factors on N flows between microorganisms and roots

Various factors may control N acquisition by roots and microorganisms and their competitive/mutualistic interactions. Unfortunately, very few studies have evaluated these effects. This hinders the statistical assessment presented in Table 4 and Figs 3–5. Therefore, we analysed individual studies that investigated biotic vs abiotic factors.

Ecosystem types – alpine or temperate grasslands, boreal or deciduous forests, and agricultural crops – feature important biotic factors that can modify interactions between roots and microorganisms for N acquisition. Note that most studies on such interactions were performed on grasslands (Table 2), and obtained results that were more reliable for such ecosystems than for forests and agricultural crops.

### 1. Root density

Root density is a key factor affecting the partitioning of available N between roots and microorganisms because of: the spatial occupation of soil volume; the high density of available C increasing microbial N immobilization; incomplete parallelism between nutrient uptake and water flux into the roots, depending on the plant’s requirement for water and nutrients and the relative availability of these resources in soil (faster and more complete water uptake with dissolved nutrients, including N); and mycorrhiza formation intensity (see the following section). In all

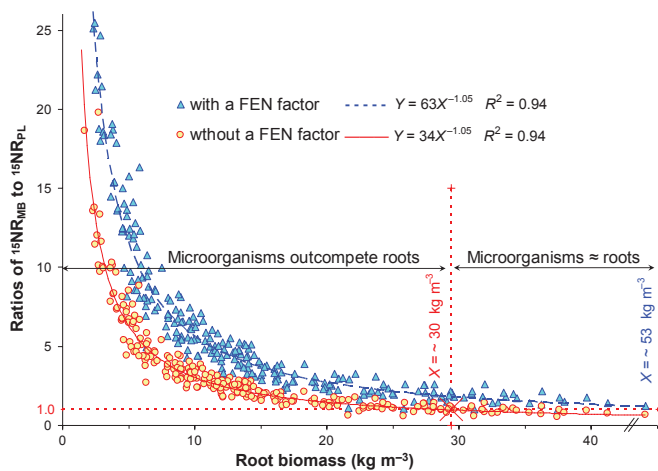


ecosystems, most roots are typically located close to the soil surface. Accordingly, the most intense competition between roots and microorganisms should be in the topsoil. However, the N content and mineralization rates are substantially higher in the topsoil, leading to greater N availability. Additionally, microbial biomass decreases with soil depth and is strongly correlated with root biomass and SOM content (Foster, 1988; Fromm *et al.*, 1993; Ekschmitt *et al.*, 2008). To our knowledge, no studies have evaluated the competition between microorganisms and roots in soils deeper than 30 cm.

In alpine grasslands, we found that the  $^{15}\text{N}$  ratio in grassland microorganisms and plants shortly after a  $^{15}\text{N}$  pulse at various soil depths is clearly correlated with root biomass (Fig. 6). This relationship allows important conclusions to be reached. Primarily, increasing root density strongly shifts the competition to much higher N uptake by roots. If root biomass is  $< c. 30 \text{ kg DW m}^{-3}$  – corresponding to  $c. 20\text{--}25 \text{ g roots per kg soil}$  – the microorganisms outcompete roots by 5–10 times. Only an extremely high root density of  $c. 53 \text{ kg DW m}^{-3}$  can enable root uptake to exceed that of microorganisms (Fig. 6).

## 2. Effect of mycorrhizas on the interactions of roots with other rhizosphere microorganisms

The roots of most plants are associated with one or more types of mycorrhiza (Allen, 1991; Newsham *et al.*, 1995; Smith & Read, 2008). Length, longevity, and branching of mycorrhizas are important in comparison with root hairs. The mutualism with



**Fig. 6** Correlation between root density and plant–microbe competition in an alpine meadow. The plant–microbe competition for nitrogen (N) was investigated up to 48 h. The dashed curve indicates plant–microbe competition for inorganic N without using an extractable factor (FEN) to correct microbial  $^{15}\text{N}$  uptake for incomplete extraction (referring to as a conservative estimate of microbial  $^{15}\text{N}$  uptake), whereas the solid curve represents plant–microbe competition for inorganic N with a factor of 0.54 to correct microbial  $^{15}\text{N}$  uptake. The value for the typical root density in this type of grasslands is  $c. 2.0 \text{ kg DW m}^{-2}$  in the upper 15-cm soil depth. Root biomass below a threshold of  $29.4 \text{ kg m}^{-3}$  implies the condition where microorganisms outcompeted roots, whereas, at a root biomass  $> 52.9 \text{ kg m}^{-3}$ , roots outcompeted microorganisms in this meadow (redrawn from Xu *et al.*, 2011b), which appears in the figure.

mycorrhizas greatly increases the active surface of roots, thereby facilitating exploration of a larger soil volume for nutrients and water uptake (Read & Perez-Moreno, 2003; Allen, 2007), and also enhancing the translocation between the roots and shoots of the host plant (Li *et al.*, 1991; Osonubi *et al.*, 1991). From the root–microorganism competition perspective, this has one important consequence: most of the C released by roots will be released via hyphal turnover and not directly from the roots to the soil (Godbold *et al.*, 2006). Therefore, promoting rhizosphere microorganisms in the presence of well-developed mycorrhizas involves the turnover of dead mycorrhizal hyphae rather than the direct release of available C by roots into the soil (Rillig *et al.*, 2001; Godbold *et al.*, 2006). Such a redistribution of easily available root-derived C alters the interactions among rhizosphere microorganisms: The large surface of mycorrhizal hyphae per gram of dried soil leads to a larger and more evenly distributed depletion of soil N (Allen, 1991; Smith & Read, 2008). Therefore, the following interactions between roots and rhizosphere microorganisms can be expected in well-developed mycorrhiza formation:

- higher belowground C allocation;
- less C release directly to soil (mostly directed to hyphae);
- higher depletion of soil N (including free amino acids and inorganic forms) and other nutrients;
- later release of available C by rapid mycorrhizal hyphal turnover;
- lower C use efficiency because of absence of available N.

We conclude that a high degree of root mycorrhiza formation is a very important factor that enhances the competitive capacity of the plants relative to free microorganisms and consequently leads to less N allocation to microorganisms (except mycorrhizal fungi). However, to our knowledge, there have been no studies that provide information on N uptake by roots, mycorrhizas, and nonmycorrhizal microorganisms, and all the studies analysed in Figs 4–6 were performed using mycorrhizal plants.

## 3. Effect of abiotic factors

The various abiotic factors affecting root–microorganism competition may be divided into two groups: factors affecting the intensity of nutrient limitation in the soil, and factors affecting the rate of soil water uptake, and consequently the rate of nutrient uptake, by roots.

The intensity of nutrient limitation can be absolute (very low N content in the soil) or relative (intensive immobilization of N by microorganisms). Very low N availability in the soil is typical, for example for alpine meadows, where microbial activities are depressed by low temperature. Nutrients are also limited in some subtropical and tropical soils because of human disturbance, for example when topsoil is eroded after the removal of vegetation. The remaining soil is then derived from the parent material, and nutrient concentrations in this soil are very low.

Very intensive N immobilization by microorganisms is usually associated with high input of plant residues with a high C : N ratio against the background of low N content in soil (Recous *et al.*, 1995). Such a situation is typical in agricultural soils after complete mineral N uptake by the harvested crop and the addition of straw. This usually significantly prolongs the mineralization of the straw

and delays N release by microorganisms (Mary *et al.*, 1996). Strong microbial N immobilization typically leads to N deficiency in crops and depresses the early development and yield of crops (Rathke *et al.*, 2006). This clearly indicates the imbalance between competition and mutualism in intensive agriculture.

Because nutrient acquisition by roots greatly depends on the delivery of nutrients to the surface of roots or mycorrhizas, the factors affecting water uptake by roots influence root N uptake (Inselsbacher & Näsholm, 2012). These factors depend directly on transpiration rates and include mainly the following factors: the intensity of photosynthetically active radiation, temperature, soil moisture, water pressure deficit, and wind velocity. The N increase attributable to all these factors also promotes the uptake of water and available N by roots before they can be taken up and used by microorganisms. Two factors – temperature and soil moisture – also directly affect microbial turnover. However, high temperature and soil moisture affect plant transpiration more than microbial turnover (Pregitzer & King, 2005). Therefore, competition is less important in warm and wet vs cold soils. This indirectly confirms the importance of initial microbial uptake of N, particularly in alpine grasslands, where transpiration is usually limited by low temperatures.

Soil pH may help determine root–microorganism competition. First, pH affects nutrient availability and changes the functions of some carriers located on the root surface (mainly in root hairs and mycorrhizal hyphae) responsible for ion uptake. Furthermore, pH greatly affects the activities and composition of microbial communities in the rhizosphere (Blagodatskaya & Anderson, 1998; Rousk *et al.*, 2010). Soil pH affects root growth and elongation (Edwards & Scott, 1974), as well as mycorrhizal colonization (van Aarle *et al.*, 2002; Read & Perez-Moreno, 2003). Accordingly, deviation of pH from the optimal (6–7) has impacts on both roots and microorganisms. The main question, however, is which of the two is suppressed more strongly. We speculate that acidification suppresses bacterial activities more strongly than those of roots: decreased soil pH therefore increases N uptake by plants. By contrast, alkalization strongly limits N uptake by roots vs microorganisms, and the latter utilize more N than at neutral soil pH. Because root N uptake can alter rhizosphere pH (Raven & Smith, 1976; Allen, 1988), the ratio between  $\text{NO}_3^-$  and  $\text{NH}_4^+$  uptake by roots could change the competition.

Competition between roots and microorganisms varies during the growth season (Jaeger *et al.*, 1999). This effect is mediated mainly by local climate and the phenological cycle of plants. These factors indirectly affect the competition via the mechanisms described above, involving temperature and precipitation and, in the rooting zone, exudate composition and root growth. In this regard, perennial grasses have an advantage in that they can acquire N early in the season: they overwinter with buds already formed and begin rapid growth as soon as the soils thaw in spring (Onipchenko *et al.*, 2009) by using stored nutrients as well as current-season N uptake (Jaeger & Monson, 1992). Additionally, plants differ in their capacity to acquire N during growth stages because the rhizosphere microbial composition changes as a result of the effects of different root exudates. We conclude that the temporal niche differentiation related to the duration of microbial and plant life

cycles contributes to the decreased competition between microorganisms and roots for N in the rhizosphere.

## VIII. Conclusions and outlook

The demand for the same nutrients leads to a significant competition between plants and microorganisms in the rhizosphere. The supply of easily available C by roots for growth and maintenance of microorganisms briefly increases this competition. Simultaneously, microbial nutrient mobilization by SOM decomposition increases, leading to an elevation in the level of mineralized nutrients available locally. This decreases the competition.

On the basis of 41  $^{15}\text{N}$  studies on plant–microbial competition for N, as well as 77 studies on the parameters of Michaelis–Menten kinetics of the uptake of nitrate, ammonium, and amino acids, we demonstrated that microorganisms show substantially faster initial uptake of all N forms. This makes them short-term winners in the competition. Conversely, the short life cycle of rhizosphere microorganisms and unidirectional N flux from soil to roots facilitates the relocation of N from microorganisms to roots. This enables plants to become winners over the long term. Thus, strong competition is weakened by mobilization of additional N from SOM, and temporal niche differentiation between microorganisms and roots.

This rapid trapping of available N by microorganisms and its subsequent slow relocation from microorganisms to plants have many ecological consequences, including maintaining ecosystem stability. First, the temporary excess of N released by litter mineralization in autumn (at periods of low N uptake by roots) allows a subsequent slow release of the N trapped by microorganisms – in spring when there is peak N demand by plants. Similarly, microbial N trapping prevents leaching from the soil or ecosystem losses. This property is probably more effective in natural ecosystems than in intensive agroecosystems, which are characterized by low microbial biomass and low available C in soils. Furthermore, plants releasing more available C in the rhizosphere are able to store more N mobilized from the SOM in microbial biomass close to their roots. This has long-term advantages for such plant species and may be beneficial from an evolutionary perspective.

We conclude that biotic factors have a strong effect on this competition, and that root density and mycorrhiza formation are crucial in shifting N acquisition to the roots.

Future studies are warranted to develop methods for quantitatively analysing the competition between roots and microorganisms for N uptake. These methods should be based on  $^{15}\text{N}$  partitioning, its redistribution dynamics after uptake, and Michaelis–Menten uptake kinetics. Although the  $K_m$  and  $V_{max}$  parameters of N uptake by roots can be estimated easily in hydroponics, the relevance of these parameters under soil conditions remains unknown. Interfering processes such as diffusion; sorption; redistribution of water and nutrients between the liquid, solid, and living phases; and competition with microorganisms may completely change root uptake compared with the theoretical predictions. We assume that these interfering processes decrease the root uptake rate and thus increase the amount of N initially stored in microorganisms.

Experimental studies along with direct and inverse modelling would help test this hypothesis.

To better evaluate the competitive abilities of plant species, particularly mutualistic stimulation of rhizosphere microorganisms, we suggest coupling the  $^{15}\text{N}$ -labelling of soil with  $^{13}\text{C}$ - and/or  $^{14}\text{C}$ -labelling of plants to investigate root exudation and rhizodeposition. This should then be related to the belowground allocated C and the N additionally mineralized from SOM.

The ecological relevance of root–microorganism competition for N suggested in this review needs to be shown by field studies that are conducted over more than a few months. This is important because the fast trapping of N by microorganisms is merely the first step of N (and probably other nutrient) retention in ecosystems. Therefore, the challenge is to investigate the second (and further) step(s) of the competition between roots and microorganisms, and to answer the question ‘How does niche differentiation in the rhizosphere contribute to ecosystem stability and productivity?’

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## References

- van Aarle IM, Olsson PA, Söderström B. 2002. Arbuscular mycorrhizal fungi respond to the substrate pH of their extraradical mycelium by altered growth and root colonization. *New Phytologist* 155: 173–182.
- Allen MF. 1991. *The ecology of mycorrhizae*. Cambridge, UK: Cambridge University Press.
- Allen MF. 2007. Mycorrhizal fungi: highways for water and nutrients in arid soils. *Vadose Zone Journal* 6: 291–297.
- Allen S. 1988. Intracellular pH regulation in plants. *ISI Atlas of Science: Animal and Plant Sciences* 1: 283–288.
- Andresen L, Jonasson S, Ström L, Michelsen A. 2008. Uptake of pulse injected nitrogen by soil microbes and mycorrhizal and non-mycorrhizal plants in a species-diverse subarctic heath ecosystem. *Plant and Soil* 313: 283–295.
- Andresen LC, Michelsen A, Jonasson S, Beier C, Ambus P. 2009. Glycine uptake in heath plants and soil microbes responds to elevated temperature,  $\text{CO}_2$  and drought. *Acta Oecologia* 35: 786–796.
- Andresen LC, Michelsen A, Jonasson S, Ström L. 2011. Seasonal changes in nitrogen availability, and root and microbial uptake of  $^{15}\text{N}$ - $^{13}\text{C}$ -phenylalanine and  $^{15}\text{N}$ -ammonium in situ at a temperate heath. *Applied Soil Ecology* 51: 94–101.
- Ashton I, Miller A, Bowman WD, Suding KN. 2008. Nitrogen preferences and plant–soil feedbacks as influenced by neighbors in the alpine tundra. *Oecologia* 156: 625–636.
- Barber SA. 1996. *Soil nutrient bioavailability: a mechanistic approach*, 2nd edn. New York, NY, USA: Wiley Press.
- Bardgett RD, Streeter TC, Bol R. 2003. Soil microbes compete effectively with plants for organic-nitrogen inputs to temperate grasslands. *Ecology* 84: 1277–1287.
- Barnard R, Barthes L, Leadley PW. 2006. Short-term uptake of  $^{15}\text{N}$  by a grass and soil micro-organisms after long-term exposure to elevated  $\text{CO}_2$ . *Plant and Soil* 280: 91–99.
- Biernath C, Fischer H, Kuzyakov Y. 2008. Root uptake of N-containing and N-free low molecular weight organic substances by maize – a  $^{14}\text{C}/^{15}\text{N}$  tracer study. *Soil Biology & Biochemistry* 40: 2237–2245.
- Blagodatskaya E, Littschwager J, Laurer M, Kuzyakov Y. 2010. Growth rates of rhizosphere microorganisms depend on competitive abilities of plants and N supply. *Plant Biosystems* 144: 408–413.
- Blagodatskaya E, Yuyukina T, Blagodatsky S, Kuzyakov Y. 2011. Turnover of soil organic matter and microbial biomass under  $\text{C}_3$ – $\text{C}_4$  vegetation change: consideration of  $^{13}\text{C}$  fractionation and preferential substrate utilization. *Soil Biology & Biochemistry* 43: 159–166.
- Blagodatskaya EV, Anderson T-H. 1998. Interactive effects of pH and substrate quality on the fungal-to-bacterial ratio and  $\text{qCO}_2$  of microbial communities in forest soils. *Soil Biology & Biochemistry* 30: 1269–1274.
- Blagodatskaya EV, Blagodatsky SA, Anderson T-H, Kuzyakov Y. 2009. Contrasting effects of glucose, living roots and maize straw on microbial growth kinetics and substrate availability in soil. *European Journal of Soil Science* 60: 186–197.
- Blagodatskaya EV, Kuzyakov Y. 2008. Mechanisms of real and apparent priming effects and their dependence on soil microbial biomass and community structure: critical review. *Biology and Fertility of Soils* 45: 115–131.
- Bloor JMG, Niboyet A, Leadley PW, Barthes L. 2009.  $\text{CO}_2$  and inorganic N supply modify competition for N between co-occurring grass plants, tree seedlings and soil microorganisms. *Soil Biology & Biochemistry* 41: 544–552.
- Bonkowski M, Villenave C, Griffiths B. 2009. Rhizosphere fauna: the functional and structural diversity of intimate interactions of soil fauna with plant roots. *Plant and Soil* 321: 213–233.
- Chen H, Marhan S, Billen N, Stahr K. 2009. Soil organic-carbon and total nitrogen stocks as affected by different land uses in Baden-Württemberg (southwest Germany). *Journal of Plant Nutrition and Soil Science* 172: 32–42.
- Cheng X, Bledsoe CS. 2004. Competition for inorganic and organic N by blue oak (*Quercus douglasii*) seedlings, an annual grass, and soil microorganisms in a pot study. *Soil Biology & Biochemistry* 36: 135–144.
- Cheng X, Bledsoe CS. 2005. Effects of annual grass senescence on  $^{15}\text{NH}_4^+$  and  $^{15}\text{N}$ -glycine uptake by blue oak (*Quercus douglasii*) seedlings and soil microorganisms in California oak woodland. *Soil Biology and Biochemistry* 37: 551–559.
- Clark FE. 1977. Internal cycling of  $^{15}\text{N}$  in shortgrass prairie. *Ecology* 73: 1148–1156.
- Clarholm M. 1985. Interactions of bacteria, protozoa and plants leading to mineralization of soil nitrogen. *Soil Biology and Biochemistry* 17: 181–187.
- Clarkson DT. 1985. Factors affecting mineral nutrient acquisition by plants. *Annual Review of Plant Physiology* 36: 77–115.
- Clemmensen KP, Sorensen PL, Michelsen A, Ström L. 2008. Site-dependent N uptake from N-form mixtures by arctic plants, soil microbes and ectomycorrhizal fungi. *Oecologia* 155: 771–783.
- Cleveland CC, Liptzin D. 2007. C:N:P stoichiometry in soil: is there a ‘Redfield ratio’ for the microbial biomass? *Biogeochemistry* 85: 235–252.
- De Graaff M-A, Classen AT, Castro HF, Schadt CW. 2010. Labile soil carbon inputs mediate the soil microbial community composition and plant residue decomposition rates. *New Phytologist* 188: 1055–1064.
- Dennis PG, Miller AJ, Hirsch PR. 2010. Are root exudates more important than other sources of rhizodeposits in structuring rhizosphere bacterial communities? *FEMS Microbiology Ecology* 72: 313–327.
- Dormaar JF. 1992. Decomposition as a process in natural grasslands. In: Couland RT, ed. *Natural grasslands, introduction and western hemisphere (ecosystems of the world)*. Amsterdam, the Netherlands: Elsevier, 121–136.
- Dunn RM, Mikola J, Bol R, Bardgett RD. 2006. Influence of microbial activity on plant–microbial competition for organic and inorganic nitrogen. *Plant and Soil* 289: 321–334.
- Edwards KL, Scott TK. 1974. Rapid growth response of corn root segments: effect of pH on elongation. *Planta* 119: 27–37.
- Ekschmitt K, Kandeler E, Poll C, Brune A, Buscot F, Friedrich M, Glerxner G, Hartmann A, Kästner M, Marhan S *et al.* 2008. Soil-carbon preservation through habitat constraints and biological limitations on decomposer activity. *Journal of Plant Nutrition and Soil Science* 171: 27–35.
- Finzi AC, Norby RJ, Calfapietra C, Gallet-Budynek A, Gielen B, Holmes WE, Hoosbeek MR, Iversen CM, Jackson RB, Kubiske ME *et al.* 2007. Increases in nitrogen uptake rather than nitrogen-use efficiency support higher rates of

- temperate forest productivity under elevated CO<sub>2</sub>. *Proceeding of National Academy of Sciences, USA* 104: 14014–14019.
- Fischer H, Ingwersen J, Kuzyakov Y. 2010. Microbial uptake of low-molecular-weight organic substances out-competes sorption in soil. *European Journal of Soil Science* 61: 504–513.
- Fischer H, Kuzyakov Y. 2010. Sorption, microbial uptake and decomposition of acetate in soil: transformations revealed by position-specific <sup>14</sup>C labeling. *Soil Biology & Biochemistry* 42: 186–192.
- Foster RC. 1988. Microenvironments of soil microorganisms. *Biology and Fertility of Soils* 6: 189–203.
- Franché C, Lindström K, Elmerich C. 2009. Nitrogen-fixing bacteria associated with leguminous and non-leguminous plants. *Plant and Soil* 321: 35–59.
- Fromm H, Winter K, Filser J, Hantschel R, Beese F. 1993. The influence of soil type and cultivation system on the spatial distributions of the soil fauna and microorganisms and their interactions. *Geoderma* 60: 109–118.
- Fusseder A, Kraus M. 1986. Individuelle Wurzelkonkurrenz und Ausnutzung der immobilen Makronährstoffe im Wurzelraum von Mais. *Flora* 178: 11–18.
- German DP, Chacon SS, Allison SD. 2011. Substrate concentration and enzyme allocation can affect rates of microbial decomposition. *Ecology* 92: 1471–1480.
- Godbold D, Hoosbeek M, Lukac M, Cotrufo M, Janssens IA, Ceulemans R, Polle A, Velthorst EJ, Scarascia-Mugnozza G, De Angelis P *et al.* 2006. Mycorrhizal hyphal turnover as a dominant process for carbon input into soil organic matter. *Plant and Soil* 281: 15–24.
- Gregory PJ. 2006. *Plant roots: growth, activity and interaction with soils*. Oxford, UK: Blackwell.
- Grogan P, Jonasson S. 2003. Controls on annual nitrogen cycling in the understory of a subarctic birch forest. *Ecology* 84: 202–218.
- Hamilton EW, Frank DA. 2001. Can plants stimulate soil microbes and their own nutrient supply? Evidence from a grazing tolerant grass. *Ecology* 82: 2397–2402.
- Hart SC, Firestone MK, Paul EA, Smith JL. 1993. Flow and fate of soil nitrogen in annual grassland and a young mixed conifer forest. *Soil Biology & Biochemistry* 25: 432–442.
- Harrison KA, Bol R, Bardgett RD. 2007. Preferences for different nitrogen forms by coexisting plant species and soil microbes. *Ecology* 88: 989–999.
- Harrison KA, Bol R, Bardgett RD. 2008. Do plant species with different growth strategies vary in their ability to compete with soil microbes for chemical forms of nitrogen? *Soil Biology & Biochemistry* 40: 228–237.
- Havlin JL, Beaton JD, Tisdal SL, Nelson WL. 1999. *Soil fertility and fertilizers. An introduction to nutrient management, 6th edn.* Bergen, NJ, USA: Prentice Hall.
- Henry HAL, Jefferies RL. 2003a. Plant amino acid uptake, soluble N turnover and microbial N capture in soils of a grazed Arctic salt marsh. *Journal of Ecology* 91: 627–636.
- Henry HAL, Jefferies RL. 2003b. Interactions in the uptake of amino acids, ammonium and nitrate ions in the Arctic salt-marsh grass, *Puccinellia phryganodes*. *Plant, Cell & Environment* 26: 419–428.
- Hill PW, Farrell M, Jones DL. 2012. Bigger may be better in soil N cycling: does rapid acquisition of small L-peptides by soil microbes dominate fluxes of protein-derived N in soil? *Soil Biology and Biochemistry* 48: 106–112.
- Hodge A, Fitter AH. 2010. Substantial nitrogen acquisition by arbuscular mycorrhizal fungi from organic material has implications for N cycling. *Proceeding of National Academy of Sciences, USA* 107: 13754–13759.
- Hodge A, Robinson D, Fitter A. 2000a. Are microorganisms more effective than plants at competing for nitrogen? *Trends in Plant Science* 5: 304–308.
- Hodge A, Stewart J, Robinson D, Griffiths BS, Fitter AH. 2000b. Competition between roots and soil microorganisms for nutrients from nitrogen-rich patches of varying complexity. *Journal of Ecology* 88: 150–164.
- Hu S, Tu C, Chen X, Gruver JB. 2006. Progressive N limitation of plant response to elevated CO<sub>2</sub>: a microbiological perspective. *Plant and Soil* 289: 47–58.
- Ingestad T, Ågren GI. 1988. Nutrient uptake and allocation at steady-state nutrition. *Physiologia Plantarum* 72: 450–459.
- Inselsbacher E, Hinko-Najera Umana N, Stange CF, Gorfer M, Schüller E, Ripka K, Zechmeister-Boltenstern S, Hood-Novotny R, Strauss J, Wanek W. 2010. Short-term competition between crop plants and soil microbes for inorganic N fertilizer. *Soil Biology and Biochemistry* 42: 360–372.
- Inselsbacher E, Näsholm T. 2012. The below-ground perspective of forest plants: soil provides mainly organic nitrogen for plants and mycorrhizal fungi. *New Phytologist* 195: 329–334.
- Jackson LE, Schimel JP, Firestone MK. 1989. Short-term partitioning of ammonium and nitrate between plants and microbes in an annual grassland. *Soil Biology & Biochemistry* 21: 409–415.
- Jaeger CH, Monson RK. 1992. Adaptive significance of nitrogen storage in *Bistorta bistortoides*, an alpine herb. *Oecologia* 92: 578–585.
- Jaeger CH, Monson RK, Fisk MC, Schmidt SK. 1999. Seasonal partitioning of nitrogen by plants and soil microorganisms in an alpine ecosystem. *Ecology* 80: 1883–1891.
- James JJ, Richards JH. 2006. Plant nitrogen capture in pulse-driven systems: interactions between root responses and soil processes. *Journal of Ecology* 94: 765–777.
- Jones DL, Hodge A, Kuzyakov Y. 2004. Plant and mycorrhizal regulation of rhizodeposition. *New Phytologist* 163: 459–480.
- Jones DL, Kemmitt SJ, Wright D, Cuttle SP, Bol R, Edwards AC. 2005. Rapid intrinsic rates of amino acid biodegradation in soils are unaffected by agricultural management strategy. *Soil Biology & Biochemistry* 37: 1267–1275.
- Jones DL, Kielland K. 2002. Soil amino acid turnover dominates the nitrogen flux in permafrost-dominated taiga forest soils. *Soil Biology & Biochemistry* 34: 209–219.
- Jungk A. 2001. Root hairs and the acquisition of plant nutrients from soil. *Journal of Plant Nutrition and Soil Science* 164: 121–129.
- Kapoor R, Mukerji KG. 2006. Rhizosphere microbial community dynamics. In: Mukerji KG, Manoharachary C, Singh J, eds. *Microbial activity in the rhizosphere*. Berlin, Germany: Springer, 55–66.
- Kastovska E, Santruckova H. 2011. Comparison of uptake of different N forms by soil microorganisms and two wet-grassland plants: a pot study. *Soil Biology and Biochemistry* 43: 1285–1291.
- Kaye JP, Hart SC. 1997. Competition for nitrogen between plants and soil microorganisms. *Trends in Ecology and Evolution* 12: 139–143.
- Kemmitt SJ, Wright D, Murphy DV, Jones DL. 2008. Regulation of amino acid biodegradation in soil as affected by depth. *Biology and Fertility of Soils* 44: 933–941.
- Kiers ET, Palmer TM, Ives AR, Bruno JF, Bruno JL. 2010. Mutualisms in a changing world: an evolutionary perspective. *Ecology Letters* 13: 1459–1474.
- Kiers ET, Rousseau RA, West SA, Denison RF. 2003. Host sanctions and the legume–rhizobium mutualism. *Nature* 425: 78–81.
- Kirchmann H, Johnston AE, Bergström LF. 2002. Possibilities for reducing nitrate leaching from agricultural land. *Ambio* 31: 404–408.
- Kuzyakov Y. 2002. Review: factors affecting rhizosphere priming effects. *Journal of Plant Nutrition and Soil Science* 165: 382–396.
- Kuzyakov Y, Domanski G. 2000. Carbon input by plants into the soil: review. *Journal of Plant Nutrition and Soil Science* 163: 421–431.
- Kuzyakov Y, Ehrensberger H, Stahr K. 2001. Carbon partitioning and below-ground translocation by *Lolium perenne*. *Soil Biology & Biochemistry* 33: 61–74.
- Kuzyakov Y, Jones DL. 2006. Glucose uptake by maize roots and its transformation in the rhizosphere. *Soil Biology and Biochemistry* 38: 851–860.
- Lal R. 2003. Global potential of soil carbon sequestration to mitigate the greenhouse effect. *Critical Reviews in Plant Sciences* 22: 151–184.
- Lambers H, Mougél C, Jaillard B, Hinsinger P. 2009. Plant–microbe–soil interactions in the rhizosphere: an evolutionary perspective. *Plant and Soil* 321: 83–115.
- LeBauer DS, Treseder KK. 2008. Nitrogen limitation of net primary productivity in terrestrial ecosystems is globally distributed. *Ecology* 89: 371–379.
- Lehmann J, Schroth G. 2003. Nutrient leaching. In: Schroth G, Sinclair EL, eds. *Trees, crops, and soil fertility: concepts and research methods*. Wallingford, UK: CAB International, 151–166.
- Li XL, Marschner H, George E. 1991. Acquisition of phosphorus and copper by VA-mycorrhizal hyphae and root-to-shoot transport in white clover. *Plant and Soil* 136: 49–57.
- Liljeroth E, Van Veen JA, Miller HJ. 1990. Assimilate translocation to the rhizosphere of two wheat lines and subsequent utilization by rhizosphere microorganisms at two soil nitrogen concentrations. *Soil Biology & Biochemistry* 22: 1015–1021.
- Lipson DA, Monson RK. 1998. Plant-microbe competition for soil amino acids in the alpine tundra: effects of freeze-thaw and dry-rewet events. *Oecologia* 113: 406–414.

- Lipson DA, Schmidt SK, Monson RK. 1999. Links between microbial population dynamics and nitrogen availability in an alpine ecosystem. *Ecology* 80: 1623–1631.
- Littschwager J, Laurer M, Blagodatskaya E, Kuzyakov Y. 2010. Nitrogen uptake and utilization as a competition factor between invasive *Duchesnea indica* and native *Fragaria vesca*. *Plant and Soil* 331: 105–114.
- Luo Y, Field CB, Jackson RB. 2006. Does nitrogen constrain carbon cycling, or does carbon input stimulate nitrogen cycling? *Ecology* 87: 3–4.
- Lynch JM. 1990. Introduction: some consequences of microbial rhizosphere competence for plant and soil. In: Lynch JM, ed. *The rhizosphere*. Chichester, UK: John Wiley & Sons Ltd, 1–10.
- Magyar G, Kun A, Oborny B, Stuefer JF. 2007. Importance of plasticity and decision-making strategies for plant resource acquisition in spatio-temporally variable environments. *New Phytologist* 174: 182–193.
- Manzoni S, Taylor P, Richter A, Porporato A, Ågren GI. 2012. Environmental and stoichiometric controls on microbial carbon-use efficiency in soils. *New Phytologist* 196: 79–91.
- Mary B, Recous S, Darwis D, Robin D. 1996. Interactions between decomposition of plant residues and nitrogen cycling in soil. *Plant and Soil* 181: 71–82.
- McFarland JW, Ruess RW, Kielland K, Pregitzer K, Hendrick R, Allen M. 2010. Cross-ecosystem comparisons of in situ plant uptake of amino acid-N and  $\text{NH}_4^+$ . *Ecosystems* 13: 177–193.
- McKane RB, Johnson LC, Shaver GR, Nadelhoffer KJ, Rastetter EB, Fry B, Giblin AE, Kjelland K, Kwiatkowski BL, Laundre JA *et al.* 2002. Resource-based niches provide a basis for plant species diversity and dominance in arctic tundra. *Nature* 415: 68–71.
- Medina-Roldán E, Bardgett RD. 2011. Plant and soil responses to defoliation: a comparative study of grass species with contrasting life history strategies. *Plant and Soil* 344: 377–388.
- Mengel K, Kirkby EA, Kosegarten H, Appel T. 2001. *Principles of plant nutrition, 5th edn.* the Netherlands: Kluwer Academic.
- Merckx R, Dijkstra A, den Hartog A, van Veen JA. 1987. Production of root derived material and associated microbial growth in soil at different nitrogen levels. *Biology and Fertility of Soils* 5: 126–132.
- Miller AE, Schimel JP, Sickman JO, Skeen K, Meixner T, Melack JM. 2009. Seasonal variation in nitrogen uptake and turnover in two high-elevation soils: mineralization responses are site-dependent. *Biogeochemistry* 93: 253–270.
- Näsholm T, Ekblad A, Nordin A, Giesler R, Höglberg M, Höglberg P. 1998. Boreal forest plants take up organic nitrogen. *Nature* 392: 914–916.
- Näsholm T, Kielland K, Ganeteg U. 2009. Uptake of organic nitrogen by plants. *New Phytologist* 182: 31–48.
- Nannipieri P, Ciardi C, Palazzi M. 1985. Plant uptake, microbial immobilization in residual soil fertilizer of urea-nitrogen in a grass legume. *Soil Science Society of American Journal* 49: 452–457.
- Newman EI, Watson A. 1977. Microbial abundance in the rhizosphere: a computer model. *Plant and Soil* 48: 17–56.
- Newsham KK, Fitter AH, Watkinson AR. 1995. Multi-functionality and biodiversity in arbuscular mycorrhizas. *Trends in Ecology and Evolution* 10: 407–411.
- Newton WE, Dilworth MJ, Sprent J, James EK. 2008. *Leguminous nitrogen-fixing symbioses*. Dordrecht, the Netherlands: Springer.
- Nguyen C. 2003. Rhizodeposition of organic C by plants: mechanisms and controls. *Agronomie* 23: 375–396.
- Nordin A, Schmidt IK, Shaver GR. 2004. Nitrogen uptake by arctic soil microbes and plants in relation to soil nitrogen supply. *Ecology* 85: 955–962.
- Oger PM, Mansouri H, Nesme X, Dessaux Y. 2004. Engineering root exudation of Lotus toward the production of two novel carbon compounds leads to the selection of distinct microbial populations in the rhizosphere. *Microbial Ecology* 47: 96–103.
- Onipchenko VG, Makarov MI, van Logtestijn RSP, Ivanov VB, Akhmetzhanova AA, Tekeev DK, Ermak AA, Salpagarova FS, Kozhevnikova AD, Cornelissen JHC. 2009. New nitrogen uptake strategy: specialized snow roots. *Ecology Letters* 12: 758–764.
- Orcutt DM, Nilsen ET. 2000. *The physiology of plants under stress: soil and biotic factors*. New York, NY, USA: John Wiley & Sons Inc.
- Osonubi O, Mlongoy K, Awotoppe OO, Atayese MO, Okali DU. 1991. Effects of ectomycorrhizal and vesicular-arbuscular mycorrhizal fungi on drought tolerance of four leguminous woody seedlings. *Plant and Soil* 136: 131–143.
- Pregitzer KS, King JS. 2005. Effects of soil temperature on nutrient uptake. In: Bassirirad H, ed. *Nutrient acquisition by plants: an ecological perspective*. Heidelberg, Germany: Springer, 277–310.
- Rasmussen J, Sauheitl L, Eriksen J, Kuzyakov Y. 2010. Plant organic N uptake is biased by inorganic C: results of triple labeling study. *Soil Biology & Biochemistry* 42: 524–527.
- Rathke G-W, Behrens T, Diepenbrock W. 2006. Integrated nitrogen management strategies to improve seed yield, oil content and nitrogen efficiency of winter oilseed rape (*Brassica napus* L.): a review. *Agriculture, Ecosystems and Environment* 117: 80–108.
- Raven JA, Smith FA. 1976. Nitrogen assimilation and transport in vascular land plants in relation to intercellular pH regulation. *New Phytologist* 76: 415–431.
- Read DJ, Perez-Moreno J. 2003. Mycorrhizas and nutrient cycling in ecosystems – a journal towards relevance? *New Phytologist* 157: 475–492.
- Recous S, Robin D, Darwis D, Mary B. 1995. Soil inorganic N availability: effect on maize residue decomposition. *Soil Biology and Biochemistry* 27: 1529–1538.
- Rillig MC, Wright SF, Nichols KA, Schmidt WF, Torn M. 2001. Large contribution of arbuscular mycorrhizal fungi to soil carbon pools in tropical forest soils. *Plant and Soil* 233: 167–177.
- Rosswall T. 1982. Microbiological regulation of the biogeochemical nitrogen cycle. *Plant and Soil* 67: 15–34.
- Rousk J, Brookes PC, Bååth E. 2010. The microbial PLFA composition as affected by pH in an arable soil. *Soil Biology & Biochemistry* 42: 516–520.
- Saharan BS, Nehra V. 2011. Plant growth promoting rhizobacteria: a critical review. *Life Science and Medicine Research* 21: 1–30.
- Schimel JP, Bennett J. 2004. Nitrogen mineralization: challenges of a changing paradigm. *Ecology* 83: 591–602.
- Schimel JP, Chapin FS III. 1996. Tundra plant uptake of amino acid and  $\text{NH}_4^+$  nitrogen in situ: plants compete well for amino acid N. *Ecology* 77: 2142–2147.
- Schimel JP, Weintraub MN. 2003. The implications of exoenzyme activity on microbial carbon and nitrogen limitation in soil: a theoretical model. *Soil Biology & Biochemistry* 35: 549–563.
- Schmidt SK, Costello EK, Nemergut DR, Cleveland CC, Reed SC, Weintraub MN, Meyer AF, Martin AM. 2007. Biogeochemical consequences of rapid microbial turnover and seasonal succession in soil. *Ecology* 88: 1379–1385.
- Simard SW, Beiler KJ, Bingham MA, Deslippe JR, Philip LJ, Teste FP. 2012. Mycorrhizal networks: mechanisms, ecology and modelling. *Fungal Biology Reviews* 26: 39–60.
- Six J, Fellwe C, Denef K, Ogle SM, Sá JCM, Albrecht A. 2002. Soil organic matter, biota and aggregation in temperate and tropical soils – effects of no-tillage. *Agronomie* 22: 755–775.
- Smith SE, Read D. 2008. *Mycorrhizal symbioses*. London, UK: Academic Press.
- Sorensen PL, Clemmensen KE, Michelsen A, Jonasson S, Ström L. 2008a. Plant and microbial uptake and allocation of organic and inorganic nitrogen related to plant growth forms and soil conditions at two subarctic tundra sites in Sweden. *Arctic and Antarctic Alpine Research* 40: 171–180.
- Sorensen PL, Michelsen A, Jonasson S. 2008b. Nitrogen uptake during one year in subarctic plant functional groups and in microbes after long-term warming and fertilization. *Ecosystems* 11: 1223–1233.
- Song MH, Xu XL, Hu QW, Tian YQ, Ouyang H, Zhou CP. 2007. Interactions of plant species mediated plant competition for inorganic nitrogen with soil microorganisms in an alpine meadow. *Plant and Soil* 297: 127–137.
- Southworth D, He X-H, Swenson W, Bledsoe CS, Horwath WR. 2005. Application of network theory of potential mycorrhizal networks. *Mycorrhiza* 15: 589–595.
- Staddon PL, Ramsey CB, Ostle N, Ineson P, Fitter AH. 2003. Rapid turnover of mycorrhizal fungi determined by AMS microanalysis of  $^{14}\text{C}$ . *Science* 300: 1138–1140.
- Talbot IM, Treseder KK. 2010. Controls over mycorrhizal uptake of organic nitrogen. *Pedobiologia* 53: 169–179.
- Templer PH, Silver WL, Pett-Ridge J, DeAngelis KM, Firestone MK. 2008. Plant and microbial controls on nitrogen retention and loss in a humid tropical forest. *Ecology* 89: 3030–3040.
- Thaler P, Pagès L. 1998. Modelling the influence of assimilate availability on root growth and architecture. *Plant and Soil* 201: 307–320.

- Tinker PB, Nye P. 2000. *Solute movement in the rhizosphere*. Oxford, UK: Oxford University Press.
- Vitousek PM, Matson PA. 1984. Mechanisms of nitrogen retention in forest ecosystems: a field experiment. *Science* 225: 51–52.
- Waisel Y, Eshel A, Kafkafi U. 2002. *Plant roots—the hidden half*, 3rd edn. New York, NY, USA and Basel, Switzerland: Marcel Dekker, Inc.
- Wanek W, Mooshammer M, Blöchl A, Hanreich A, Richter A. 2010. Determination of gross rates of amino acid production and immobilization in decomposing leaf litter by a novel  $^{15}\text{N}$  isotope pool dilution technique. *Soil Biology and Biochemistry* 42: 1293–1302.
- Whiteside MD, Digman MA, Gratton E, Treseder KK. 2012. Organic nitrogen uptake by arbuscular fungi in a boreal forest. *Soil Biology and Biochemistry* 55: 7–13.
- Wichern F, Eberhardt E, Mayer J, Joergensen RG, Muller T. 2008. Nitrogen rhizodeposition in agricultural crops: methods, estimates and future prospects. *Soil Biology & Biochemistry* 40: 30–48.
- Wu H, Dannenmann M, Fanselow N, Wolf B, Yao Z, Wu X, Brüggemann N, Zheng XH, Han XG, Ditter K *et al.* 2011. Feedback of grazing on gross rates of N mineralization and inorganic N partitioning in steppe soils of Inner Mongolia. *Plant and Soil* 340: 127–139.
- Xu XL, Kuzyakov Y, Stange F, Richter A, Wanek W. 2008. Light affected the competition for inorganic and organic nitrogen between maize and soil microorganisms. *Plant and Soil* 304: 59–72.
- Xu XL, Ouyang H, Cao GM, Richter A, Wanek W, Kuzyakov Y. 2011a. Dominant plant species shift their nitrogen uptake patterns in response to nutrient enrichment caused by a fungal fairy in an alpine meadow. *Plant and Soil* 341: 495–504.
- Xu XL, Ouyang H, Pei ZY, Zhou CP. 2003. The fate of short-term  $^{15}\text{N}$  labeled nitrate and ammonium added to an alpine meadow in the Qinghai-Xizang Plateau, China. *Acta Botanica Sinica* 45: 276–281.
- Xu XL, Ouyang H, Pei ZY, Zhou CP. 2004. Long-term partitioning of  $^{15}\text{N}$  labeled ammonium and nitrate among different components in an alpine meadow ecosystem. *Acta Botanica Sinica* 46: 279–283.
- Xu XL, Ouyang H, Richter A, Wanek W, Cao GM, Kuzyakov Y. 2011b. Spatio-temporal patterns of plant-microbial competition for inorganic nitrogen in an alpine meadow. *Journal of Ecology* 99: 563–571.
- Zak DR, Pregitzer KS, King JS, Holmes WE. 2000. Elevated atmospheric  $\text{CO}_2$ , fine roots and the response of soil microorganisms: a review and hypothesis. *New Phytologist* 147: 201–222.
- Zak DR, Groffman PM, Pregitzer KS, Christensen S, Tiedje JM. 1990. The vernal dam: plant-microbe competition for nitrogen in northern hardwood forests. *Ecology* 71: 651–656.
- Zogg GP, Zak DR, Pregitzer KS, Bruton AJ. 2000. Microbial immobilization and the retention of anthropogenic nitrogen in a northern hardwood forest. *Ecology* 81: 1858–1866.

## Supporting Information

Additional supporting information may be found in the online version of this article.

**Table S1** Michaelis–Menten kinetics of nitrogen (N) uptake from  $\text{NO}_3^-$ ,  $\text{NH}_4^+$ , and various amino acids by microorganisms; source publication for data in Table 3 and Fig. 3

**Table S2** Michaelis–Menten kinetics of nitrogen (N) uptake from  $\text{NO}_3^-$ ,  $\text{NH}_4^+$ , and various amino acids by plants; source publication for data in Table 3 and Fig. 3

**Table S3** Publications of  $^{15}\text{N}$ -labelling experiments used in this study to evaluate the competition for nitrogen (N) between plants and microorganisms

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